

PEARL

Are We Speaking the Same Language? Recommendations for a Definition and Categorization Framework for Plastic Debris

Hartmann, Nanna B.; Hüffer, Thorsten; Thompson, Richard C.; Hassellöv, Martin; Verschoor, Anja; Daugaard, Anders E.; Rist, Sinja; Karlsson, Therese; Brennholt, Nicole; Cole, Matthew; Herrling, Maria P.; Hess, Maren C.; Ivleva, Natalia P.; Lusher, Amy L.; Wagner, Martin

Published in:

Environmental Science & Technology

DOI:

[10.1021/acs.est.8b05297](https://doi.org/10.1021/acs.est.8b05297)

Publication date:

2019

Link:

[Link to publication in PEARL](#)

Citation for published version (APA):

Hartmann, N. B., Hüffer, T., Thompson, R. C., Hassellöv, M., Verschoor, A., Daugaard, A. E., Rist, S., Karlsson, T., Brennholt, N., Cole, M., Herrling, M. P., Hess, M. C., Ivleva, N. P., Lusher, A. L., & Wagner, M. (2019). Are We Speaking the Same Language? Recommendations for a Definition and Categorization Framework for Plastic Debris. *Environmental Science & Technology*, 53(3), 1039-1047. <https://doi.org/10.1021/acs.est.8b05297>

All content in PEARL is protected by copyright law. Author manuscripts are made available in accordance with publisher policies. Wherever possible please cite the published version using the details provided on the item record or document. In the absence of an open licence (e.g. Creative Commons), permissions for further reuse of content should be sought from the publisher or author.

Download date: 16. Sep. 2024

1 Disclaimer

2 'This is a copy of the accepted paper as submitted for publication. Readers are advised to
3 refer to the final version of the paper which can be found at

4

5 <https://pubs.acs.org/doi/10.1021/acs.est.8b05297>

6 **Are we speaking the same language? Recommendations for a definition and**
7 **categorization framework for plastic debris**

8

9 **Nanna B. Hartmann^{1,*}, Thorsten Hüffer^{2,*}, Richard Thompson³, Martin Hassellöv⁴, Anja**
10 **Verschoor⁵, Anders Egede Daugaard⁶, Sinja Rist¹, Therese Karlsson⁴, Nicole**
11 **Brennholt⁷, Matthew Cole⁸, Maria P. Herrling⁹, Maren Heß¹⁰, Natalia P. Ivleva¹¹, Amy L.**
12 **Lusher¹², Martin Wagner^{13,*}**

13 ¹ Technical University of Denmark, Department of Environmental Engineering,
14 Bygningstorvet B115, 2800 Kgs. Lyngby, Denmark

15 ² University of Vienna, Department of Environmental Geosciences; Environmental Science
16 Research Network; and Research Platform Plastics in the Environment and Society
17 (PLENTY), Althanstrasse 14, 1090 Vienna, Austria

18 ³ School of Biological and Marine Sciences, Plymouth University, Plymouth PL4 8AA, United
19 Kingdom

20 ⁴ University of Gothenburg, Department of Marine Sciences, Kristineberg 566, 45178
21 Fiskebäckskil, Sweden

22 ⁵ National Institute for Public Health and the Environment, Antonie van Leeuwenhoeklaan 9,
23 3721 MA Bilthoven, The Netherlands

24 ⁶ Technical University of Denmark, Department of Chemical and Biochemical Engineering,
25 Danish Polymer Centre, Søtofts Plads B227, 2800 Kgs. Lyngby, Denmark

26 ⁷ Federal Institute of Hydrology, Department Biochemistry and Ecotoxicology, Am Mainzer
27 Tor 1, 56068 Koblenz, Germany

28 ⁸ Marine Ecology & Biodiversity, Plymouth Marine Laboratory, Prospect Place, The Hoe,
29 Plymouth PL1 3DH, United Kingdom

30 ⁹ Ovivo Switzerland AG, Hauptstrasse 192, 4147 Aesch, Switzerland

31 ¹⁰ North Rhine Westphalia State Agency for Nature, Environment and Consumer Protection,
32 Department Water Management, Water Protection, Postfach 101052, 45610
33 Recklinghausen, Germany

34 ¹¹ Technical University of Munich, Institute of Hydrochemistry, Chair of Analytical Chemistry
35 and Water Chemistry, Marchioninstr. 17, 81377 Munich, Germany

36 ¹² Norwegian Institute for Water Research (NIVA), 0349 Oslo, Norway

37 ¹³ Norwegian University of Science and Technology (NTNU), Department of Biology, 7491
38 Trondheim, Norway

39 * Corresponding authors: nibh@env.dtu.dk, thorsten.hueffer@univie.ac.at,
40 martin.wagner@ntnu.no

41

42 **Keywords:** harmonization, legislation, litter, microplastics, plastics, terminology

43

44 **Synopsis**

45 Plastic pollution is a global issue. However, there is no consensus on how to define and
46 categorize plastic debris, for instance in terms of materials or size classes. As this ambiguity
47 creates miscommunication, we propose a framework to define plastic debris based on

48 material properties and categorize it according to size, shape, color, and origin. This should
49 help to clarify what we actually mean when we talk about plastic debris.

50 **Abstract**

51 The accumulation of plastic litter in natural environments is a global issue. Concerns over
52 potential negative impacts on the economy, wildlife, and human health provide strong
53 incentives for improving the sustainable use of plastics. Despite the many voices raised on
54 the issue, we lack a consensus on how to define and categorize plastic debris. This is
55 evident for microplastics, where inconsistent size classes are used, and where the materials
56 to be included are under debate. While this is inherent in an emerging research field, an
57 ambiguous terminology results in confusion and miscommunication that may compromise
58 progress in research and mitigation measures.

59 Therefore, we need to be explicit on what exactly we consider plastic debris. Thus, we
60 critically discuss the advantages and disadvantages of a unified terminology, propose a
61 definition and categorization framework and highlight areas of uncertainty.

62 Going beyond size classes, our framework includes physico-chemical properties (polymer
63 composition, solid state, solubility) as defining criteria and size, shape, color, and origin as
64 classifiers for categorization. Acknowledging the rapid evolution of our knowledge on plastic
65 pollution, our framework will promote consensus-building within the scientific and regulatory
66 community based on a solid scientific foundation.

67 1 Introduction

68 Plastic pollution is a substantial environmental problem. Plastic debris, that is, plastic items
69 occurring in natural environments without fulfilling an intended function, is persistent, mobile,
70 and ubiquitous in terrestrial and aquatic environments, including urban, rural, and remote
71 locations. Large plastic litter is readily visible and adversely affects wildlife species through
72 entanglement, ingestion and lacerations.¹ Microscopic plastic debris (i.e., microplastics) has,
73 until recently, largely been an overlooked part of plastic pollution. This has changed in the
74 last decade, over which time growing scientific, public, and political interest has focused on
75 the smaller size fractions, in particular those in the micrometer size range.² Today, research
76 into the formation, features, further fragmentation, chemical interactions, environmental fate,
77 and potential impacts of microplastics is increasingly abundant.³

78 The term 'microlitter' was used in 2003 to describe the fine fraction of marine plastic litter with
79 sizes of 63–500 μm .³ Similarly, mesolitter, macrolitter, and megalitter were defined as having
80 sizes of <5 to 10 mm, <10 to 15 cm or measuring decimeters or more across, respectively.⁴

81 In 2004, the term 'microplastics' was popularized to describe truly microscopic plastic
82 fragments with typical diameters down to $\sim 20 \mu\text{m}$.⁵ While this paper described the

83 accumulation of microplastics in the seas around the United Kingdom, it did not define them.

84 In 2008, experts attending a meeting hosted by NOAA proposed a working definition in which
85 microplastics are all plastic particles <5 mm in diameter,⁶ which has become the most

86 frequently used definition. Although not yet detected in environmental samples, sub-micron
87 sized particles are expected to form in the environment through fragmentation of larger

88 plastics.^{7, 8} These have been termed nanoplastics.^{9, 10} Due to the evolving research on plastic

89 debris, a certain nomenclature has developed. Nonetheless, the terminology remains

90 ambiguous and conflicting, for instance regarding the size classes (Figure 1). So far, "[t]here

91 *is no internationally agreed definition of the size below which a small piece of plastic should*

92 *be called a microplastic*".¹¹

93 In the context of this paper, categorization refers to a systematic division of plastic debris into
94 groups according to pre-established criteria. Hereby, plastic objects are grouped based on
95 similarity. A commonly used categorization system is based on size using the prefixes of
96 mega-, macro-, meso-, micro-, and nano. In addition, plastic debris can also be categorized
97 based on their origin, shape, and polymer type.¹² However, a systematic framework for
98 categorizing plastic debris is currently missing.

99 While most of the discourse on what makes a plastic item, for instance, a “microplastic”
100 focuses on size as only criterion,¹³ we first need to revisit the question of what plastics
101 actually are. This is important because – apart from the commodity polymers – there is no
102 consensus on which materials to include in the term ‘plastics’. For instance, some studies
103 consider cellophane, i.e., regenerated cellulose, as plastics^{14, 15} while it can be argued that it
104 is not. In addition, definition criteria from polymer sciences are not stringently applicable to
105 plastic debris. For instance, rubber is not plastic according to some polymer chemistry
106 definitions.¹⁶ Yet, environmental researchers consider rubber-containing tire wear a major
107 component of microplastic pollution.^{17, 18} The same is true for paint particles. To clarify, we
108 discuss basic physico-chemical properties as ‘definition criteria’ before considering size,
109 shape, color, and origin as ‘classification criteria’ for the categorization of plastic debris.

110 **2 Do we need a common terminology?**

111 The lack of consensus on a definition and categorization of plastic debris results in an
112 ambiguous communication and the generation of incomparable data. While this situation
113 inevitably calls for a harmonization, we need to keep in mind the implications of such a
114 framework. Categorizing plastic debris into different classes (e.g., sizes) implicitly suggests
115 that the items within one category have some 'likeness' whereas plastics in different
116 categories are somehow different. This may be perceived as similarity in hazardous
117 properties or environmental behavior. Such connotation has emerged for the term
118 microplastics, using size as a key feature, already.¹⁹ On the downside, this may point
119 research towards properties that are irrelevant and result in neglecting features that are
120 potentially important. A framework can, thus, shape the research field and affect current and
121 future mitigation measures based on how it frames the problem. This will also affect the risk
122 perception and the hypotheses generated to examine it.²⁰

123 In the area of engineered nanomaterials, the process of agreeing on a common terminology
124 has been ongoing for more than a decade and is under continuous debate^{21, 22} and
125 revision.²³ For nanomaterials, the European Commission 'Recommendation on a Definition
126 of Nanomaterials' states that: "*an upper limit of 100 nm is commonly used by general
127 consensus, but there is no scientific evidence to support the appropriateness of this value.*"²⁴
128 It has been further specified that "*clear [size] boundaries were primarily introduced with the
129 regulatory purpose of the definition in mind rather than for scientific reasons.*"²³ Hence, the
130 size boundaries are not scientifically justified but rather based on pragmatic reasons and
131 general consensus. As behavior and toxicity will also depend on properties other than size, a
132 purely scientific definition of nanomaterials may never be achieved – at least not if it shall
133 have any practical value.

134 For plastic debris, similar considerations do apply: There is no clear scientific justification for
135 the currently applied size boundaries. The 5 mm upper limit for microplastics proposed by
136 NOAA⁶ is somewhat biologically informed as particles of this size were considered more

137 likely to be ingested compared to larger items. Still, the decision on size limits is not based
138 on actual evidence but rather on pragmatism.¹¹

139 Ultimately, the question whether to establish a definition/categorization framework for plastic
140 debris is at the heart of two conflicting points of view. On the one hand, there is the notion to
141 refute any attempt to unify the terminology as this restricts scientific freedom and narrows
142 down the scientific focus to what is included in the definition. On the other hand, there is the
143 view that a globally accepted definition is an essential prerequisite to tackle the issue,
144 especially from a regulatory perspective.²⁵ As environmental scientists, we work in the space
145 between these poles and can neither ignore the importance of academic freedom nor our
146 obligation to support science-based policy-making. While we acknowledge that a flexible,
147 adaptive, and continuously updated framework would be ideal for science, we recognize that
148 this conflicts with regulatory needs and processes. For instance, the control of microplastic
149 emissions will depend on a common definition. Accordingly, the discourse needs to focus on
150 developing a pragmatic and workable framework enabling effective regulation while not
151 restraining scientific freedom.

152 **3 Guiding principles for formulating a definition/categorization framework**

153 When developing a framework for defining and categorizing plastic debris, we considered the
154 following guiding principles, assumptions and disclaimers:

- 155 1) A definition/categorization framework should not be tied to current methodological and
156 analytical capabilities as these evolve constantly.
- 157 2) A definition/categorization framework should not be limited to size as sole criterion as
158 properties other than size contribute to the impacts of plastic debris.
- 159 3) A definition can be based on scientific criteria using the physical and chemical properties
160 of the materials included as plastics.
- 161 4) A categorization cannot be purely science-based because the biologically relevant
162 properties needed to categorize plastic objects are not well understood.
- 163 5) Accordingly, any categorization will, to some extent, be arbitrary and must be based on
164 conventions formed by consensus and guided by pragmatism.
- 165 6) Thus, the proposed definition/categorization framework is a recommendation that aims at
166 promoting consensus-building on a common terminology.
- 167 7) Consensus-building in academia is a dynamic process rather than a one-time decision.
168 Thus, the proposed framework must be subjected to criticism and revision.
- 169 8) Regardless of the existence of this or any other definition/categorization framework,
170 scientific data should always be reported in the most comprehensive way, that is, in
171 accordance with the latest state of the science.
- 172 9) A material should not be excluded from the framework based on its degradability or state
173 of degradation as even “degradable” materials will form smaller fragments before they
174 mineralize.
- 175 10) The main audience of this framework are researchers, as a common terminology needs
176 to form in the community producing the primary knowledge on plastic pollution. However,
177 the framework can also serve as point of departure for policy-makers and the regulatory
178 community.

179 **4 Recommendation of a definition/categorization framework for plastic debris**

180 To structure the discussion on what plastic debris is, we propose a framework which
181 differentiates between defining criteria that address basic properties and auxiliary criteria for
182 categorizing plastic debris (principle 3 and 4, Figure 2). According to Merriam-Webster's
183 dictionary, a definition is "*a statement expressing the essential nature of something.*" A good
184 question to ask about any definition is therefore: does it actually capture the property that we
185 are trying to define? For plastic debris, we consider the following as relevant defining
186 properties: chemical composition, solid state, and solubility (criteria I-III). These will
187 determine whether a material classifies as 'plastic' and, thus, 'plastic debris' when found in
188 natural environments. For further categorization, we discuss size, shape and structure, color,
189 and origin (criteria IV-VII).

190

191 **4.1 Criterion I: Chemical composition**

192 The chemical composition is the most fundamental criterion for defining plastic debris. Some
193 disagreement exists on which polymers should be considered 'plastics'. For instance,
194 according to ISO plastic is a "*material which contains as an essential ingredient a high*
195 *molecular weight polymer and which, at some stage in its processing into finished products,*
196 *can be shaped by flow.*"¹⁶ In contrast to thermoplastics and thermosets, some elastomers
197 (e.g., rubbers) are excluded from this definition. This mirrors the industrial landscape and,
198 thus, has historic rather than scientific reasons. Questions, therefore, arise whether materials
199 derived from rubber or inorganic/hybrid polymers (e.g., silicone) qualify as plastics. Also, are
200 plastics with a high content of low-molecular weight additives (e.g., polyvinyl chloride (PVC)
201 containing >50% plasticizers) included? And should polymer composites fall under such a
202 definition? Finally, should crystalline fibers, which are not shaped by flow, be excluded from a
203 definition even though they are composed of the same polymers as other plastic debris?
204 These questions reflect the different perspectives of material and environmental sciences.

205 *a. Polymers*

206 As the ISO definition of plastics excludes certain materials, which are relevant in
207 environmental terms (e.g., elastomers), we use a broader definition as point of departure.
208 IUPAC defines a polymer as a “*molecule of high relative molecular mass, the structure of*
209 *which essentially comprises the multiple repetition of units derived, actually or conceptually,*
210 *from molecules of low relative molecular mass.*”²⁶ Typically, polymers have a molecular mass
211 of $>10,000 \text{ g mol}^{-1}$.²⁷

212 As a next level, we can use the origin of the polymer as criterion and differentiate between
213 natural and artificial (man-made, synthetic) polymers. With regard to the former, there is
214 agreement that natural polymers (e.g., DNA, proteins, wool, silk, cellulose) are not plastics
215 while synthetic polymers commonly are. Modified natural polymers, natural rubber and
216 cellulose further processed to make the final polymer (rayon and cellophane) for instance,
217 represent a special case. Because these polymers are heavily modified, they can also be
218 considered artificial and should be included in a definition of plastic debris.

219 The inclusion of natural polymers that have been slightly processed (e.g., dyed wool) is more
220 difficult. This predominantly concerns polymer fibers used for textiles and we do not have
221 sufficient information to benchmark the occurrence and impacts of natural, modified natural,
222 and synthetic fibers, respectively. However, because their essential ingredient is a natural
223 polymer, we propose to exclude slightly modified natural fibers from a definition.²⁸

224 Conventional plastics are petroleum-based and include the commodity plastics polyethylene
225 (PE), polypropylene (PP), polyurethane, polyethylene terephthalate (PET), polystyrene (PS),
226 and PVC. Recently, bio-based plastics synthesized from non-fossil feedstock have entered
227 the market. Bio-based monomers can be used to make the conventional polymers (e.g., bio-
228 PET, bio-PE) or biodegradable polymers such as polylactic acid and
229 polyhydroxyalkanoates.²⁹ A third type of plastics is mainly produced from inorganic
230 monomers. These inorganic or hybrid polymers – silicone is the most prominent example –
231 are usually excluded from plastics definitions, since they are elastomers. However, because
232 all three polymer classes are synthetic and are emitted to the environment, we recommend
233 including them in a definition of plastic debris.

234 *b. Additives*

235 Plastics can contain a broad range of low molecular weight additives to improve their
236 processability, properties, and performance. They are, thus, an essential part of the
237 formulation. The major classes of additives include plasticizers, stabilizers, flame retardants,
238 flow modifiers, processing aids, impact modifiers, and antioxidants.^{30, 31} In addition, pigments,
239 biocides, and fragrances can be added. Additives and other small molecules present in
240 plastics (e.g., monomer residues or by-products formed during production) may be
241 toxicologically relevant when leaching from the material. Nonetheless, they are not of specific
242 importance for a definition because the polymer backbone, not its additive content, defines a
243 plastic material. Polymers containing high amounts of additives (e.g., PVC) represent a
244 special case. According to REACH,³² substances with an additive content of >50% are not
245 polymers. In contrast, we propose to exclude the additive content as criterion because it will
246 change continuously after the release into in the environment.

247 *c. Copolymers*

248 Some synthetic polymers are produced “*from more than one species of monomer.*”²⁶ These
249 include copolymers of acrylonitrile-butadiene-styrene (ABS), ethylene-vinyl acetate (EVA),
250 and styrene-butadiene rubber (SBR). ABS and EVA are thermoplastic polymers (i.e.,
251 ‘plastics’ according to ISO) and, thus, can be considered plastic debris when found in the
252 environment. The same argument can be applied to thermoplastic elastomers, such as
253 styrenic block copolymers, thermoplastic olefins, and thermoplastic polyurethanes, which are
254 widely used in automotive manufacturing. In line with the arguments made above, SBR (also
255 an elastomer) and other synthetic rubber copolymers should be included in a definition.

256 *d. Composites*

257 Polymer composites consist of at least two components; the polymer matrix and
258 (non)polymeric reinforcement. Classical thermoset composites include glass fiber-reinforced
259 polyester or graphite reinforced epoxy, both used for instance for boat hulls. This also
260 includes thermoplastics filled with various inorganic materials to reduce costs or improve

261 properties. Likewise, polyester textiles are often mixed with cotton or wool. We recommend
262 including composites into a definition of plastic debris because synthetic polymers are an
263 essential ingredient. However, it remains unknown whether setting a minimum polymer
264 content of a material to qualify as plastics is appropriate and feasible.

265 *Special cases 1: Surface coatings*

266 One special case of composites are paint particles found in the environment. Surface
267 coatings (such as paints) are applied as a thin layer to a surface for aesthetic or protective
268 reasons.³³ Coatings are formulated, multi-component systems consisting of binders,
269 pigments, fillers and extenders, solvents, and additives. Polymers are used as film formers
270 and include (modified) natural resins, curing coating systems (e.g., polyester, alkyds, epoxy
271 resin, urethane resins), and physically drying systems (acryl and vinyl (co)polymers).³³

272 The central question for including coating particles in a definition is whether the synthetic
273 polymers used in surface coatings are considered plastics. Recent government reports argue
274 that they should.^{34, 35} Indeed, particles originating from dried paints and lacquers containing
275 cured thermosets can be considered plastic debris. Examples are coatings based on
276 polyesters, vinyl esters, polyurethanes as well as epoxy, phenolic, acrylic resins and alkyd.³⁴
277 Accordingly, particles derived from paints and surface coatings containing synthetic polymers
278 as an essential ingredient should be included in a definition. However, as in the case of
279 composites, setting a threshold for a minimal polymer content is currently not possible.

280 *Special case 2: Tire wear particles*

281 Driving vehicles releases particles due to the abrasion of tires, termed tire wear particles
282 (TWP). Some agencies have considered TWP to be 'microplastics'^{34, 36-38} because tires
283 usually contain 40–60% of synthetic polymers (e.g., SBR or polybutadiene rubber). The
284 exact composition of tires depends on their application.¹⁸ To classify TWP as plastic debris,
285 two questions need to be addressed: First, are rubbers plastics? Here, we argue that they
286 should be covered by the proposed definition (see criterion Ic). Second, do we need to take
287 into account a changing chemical and material composition during weathering? As an

288 example, TWP will aggregate with road particles and form tire wear and road particles
289 (TWRP) with a lower total polymer content. We argue it is not feasible to determine the
290 polymer content of TWRP as this would need to happen for each individual particle. This is
291 also true for other plastic particles forming heteroaggregates with other particulate matter.
292 Accordingly, we propose to refer to the original material and to include TWP/TWRP in the
293 definition because synthetic polymers are an essential ingredient of tires.

294

295 **4.2 Criterion II: Solid state**

296 While it might be common sense that plastics are solid materials, some polymers can be
297 wax-like, semi-solid or liquid. According to the Global Harmonized System for Classification
298 and Labelling of Chemicals (GHS) a solid substance or mixture “*does not meet the*
299 *definitions of liquid or gas.*” As most polymers have a vapor pressure of <300 kPa (at 50 °C)
300 and an initial melting point of >20 °C (T_m at 101.3 kPa) they are solid.³⁹ For most materials,
301 the T_m determines the difference between the solid and the liquid state. However, amorphous
302 and semi-crystalline plastics will behave differently when heated. Amorphous polymers (e.g.,
303 polymethyl methacrylate, ABS, PS) are hard, brittle materials below their glass transition
304 temperature (T_g), whereas they become viscous and free flowing above. Semi-crystalline
305 polymers (e.g., polyamide, polycarbonate, PE, PET, PP, PVC) have both, a T_g as well as a
306 T_m . These polymers will be hard and brittle below their T_g but ductile, soft, and form stable
307 below their T_m , and liquid above.

308 Plastics are used both as hard and brittle as well as softer and more ductile materials
309 (plasticized PVC, PE, PP) and depending on molecular weight exist as waxy, semi-solids
310 over a broad temperature range. For some polymers (e.g., rubber, PE, PP, PVC), T_g is
311 relatively low. Accordingly, they are soft solids at ambient temperatures. Nevertheless, semi-
312 crystalline polymers have a T_m high enough to classify them as solid according to GHS and
313 can be included in a definition of plastic debris.

314 In contrast, amorphous polymers lack a specific T_m . Therefore, we propose to consider the T_g
315 as a defining value. Accordingly, amorphous polymers with a $T_g > 20$ °C should be included in
316 a definition. Here, the properties of the bulk materials should be considered. However, the
317 question remains whether wax-like polymers with a $T_g < 20$ °C should be included as well. In
318 this regard, the combination with other physico-chemical properties, such as viscosity,
319 modulus of elasticity or tension at constant elongation, might be helpful.

320 *Special case 3: Polymer gels*

321 Polymer gels are often perceived as liquid rather than solid, due to their high liquid content
322 and their soft and flexible appearance. However, in macromolecular science a gel is indeed
323 “a solid composed of at least two components, one of which (polymer) forms a three-
324 dimensional network [...] in the medium of the other component (liquid).”⁴⁰ Polymer gels
325 come from a natural (e.g., gelatin, agarose) or synthetic feedstock (polyacrylamide, polyvinyl
326 alcohol (PVA), low molecular weight polyethylene glycol (PEG)) and are used in a wide
327 variety of applications. For instance, polyacrylamide copolymers are used as flocculation
328 agents during wastewater treatment. While these gels are “solid” from a chemical
329 perspective, they will become soft and viscous in water. Although this does not make them
330 benign *per se* (we simply do not know), we argue that polymer gels are not particulate matter
331 once in aquatic environments and should, therefore, be excluded from a definition and
332 treated as an independent category of environmental polymers.

333

334 **4.3 Criterion III: Solubility**

335 Another important aspect is the polymer’s solubility. Most conventional polymers are poorly
336 soluble in water, but some synthetic polymers readily dissolve in water (e.g., PVA or low
337 molecular weight PEG). We propose using solubility as a criterion to define plastic debris and
338 apply the REACH guidance provided by ECHA. Here, a substance is considered poorly
339 soluble if their water solubility is < 1 mg L⁻¹ at 20 °C.⁴¹ Polymers that are poorly soluble
340 according to REACH should be included in a definition of plastic debris.

341

342 **4.4 Criterion IV: Size**

343 Size is the criterion most frequently used to categorize plastic debris, with size classes
344 typically attributed with the nomenclature of nano-, micro-, meso-, and macroplastics. Particle
345 size will be of major ecological relevance because it is one important factor determining the
346 item's interaction with biota and its environmental fate.⁴²⁻⁴⁴ Currently, there is no clear
347 consensus on the use of size categories (Figure 1). Often, size limits are operationally
348 defined by the sampling method. As an example, some authors set the lower size limit of
349 microplastics to 333 μm because a 333 μm mesh plankton net is used for sampling.⁶

350 From a nomenclature point of view, it is intuitive to categorize the plastics based on the
351 conventional units of size. Accordingly, plastics with sizes in the nanometer scale (1–1,000
352 nm) should be nanoplastics. Following this reasoning and using the SI prefixes for length,
353 microplastics would have sizes of 1–1,000 μm , followed by milliplastics (1–10 mm),
354 centiplastics (1–10 cm), deciplastics (1–10 dm). This, however, conflicts with the current
355 terminology. For example, nanoplastics and microplastics are typically considered to be 1–
356 100 nm and 1–5,000 μm in size, respectively.⁴⁵ Accordingly, new size categories, fully
357 consistent with the SI nomenclature, would have little chance of being adopted by the
358 scientific community. As a pragmatic compromise, we propose the following categories:

- 359 - Nanoplastics: 1 to <1,000 nm,*
- 360 - Microplastics: 1 to <1,000 μm ,
- 361 - Mesoplastics: 1 to <10 mm,
- 362 - Macroplastics: 1 cm and larger.

363 *To conform to existing definitions of nanomaterials, a sub-division in nanoplastics (1 to <100
364 nm) and submicron-plastics (100 to <1,000 nm) can be made.

365 Another important question relates to the dimensions of the plastic item. Is it sufficient that it
366 possesses the given size in one, two or three dimensions to fall into one of the categories?

367 Current size classes for microplastics refer to the largest dimension of the item. This is
368 straightforward for relatively spherical particles but more ambiguous for irregular particles
369 and fibers.⁴⁵ For example, should a fiber with a diameter (i.e., two dimensions) of 500 µm
370 and a length (i.e., one dimension) of 20 mm be classified as microplastic or macroplastic? If
371 two dimensions in the micrometer range would be sufficient to qualify as a microplastic, this
372 would theoretically imply that a thin thread of infinite length would still be a microplastic. This
373 would correspond to the current practice of determining size by filtration through a net of a
374 certain mesh size or by microscopy, whereby two dimensions are considered. However, a
375 classification should not be based on current practices, which may change as the
376 methodology advances. We, therefore, propose to use the largest dimension as classifier for
377 the size category. The rationale behind this is that the largest dimension of an item will
378 mainly determine the ingestion by biota. For fibers, we do recognize that the diameter may
379 be more relevant and suggest that the dimensions used for categorization should then be
380 defined in the specific study.

381

382 **4.5 Criterion V: Shape and structure**

383 Aside from size, plastic debris is commonly categorized based on shape, structure, and
384 color. Frequent descriptors of shape are: spheres, beads, pellets, foams, fibers, fragments,
385 films, and flakes.⁴⁶⁻⁴⁹ These are worth revisiting in order to apply a more stringent
386 classification. The first three (spheres, pellets, beads) are often used synonymously.
387 Additionally, the terms 'beads' and 'pellets' hints towards the origin of the particles, such as
388 microbeads in cosmetics and pre-production pellets used for plastic manufacturing. If the
389 origin of the specific particle can indeed be elucidated this would be an appropriate
390 terminology. However, as this is often challenging it is instead beneficial to adopt more
391 neutral descriptors, such as 'spheres' for particles with every point on its surface having the
392 same distance from its center. The terms 'spheroids' and 'cylindrical pellets' can be used for
393 approximate spheres and cylindrical shapes, respectively.

394 'Fragments' also represent a rather ambiguous category. It is commonly adopted to describe
395 particles with irregular shape. The term, however, implies that these have been formed by
396 fragmentation in the environment, which is not necessarily the case. For instance, irregular
397 abrasives used in cosmetics are produced as such⁵⁰⁻⁵² and cannot be distinguished from
398 particles generated by secondary fragmentation. While the category 'fragment' is likely to
399 persist in the literature, an alternative and more accurate term is 'irregular particles'.

400 The category of 'films' is rather straight-forward as this includes planar objects which are
401 considerably smaller in one dimension than in the other two. It is useful to classify films
402 separately and it is often feasible to make that distinction for items >300 µm. Smaller objects
403 tend to overlap and, due to practical constraints, may be pooled with 'irregular particles'.

404 Plastics that are significantly longer in one than wide in two dimensions (length-to-diameter
405 ratio) are commonly (and interchangeably) described as fibers or filaments, with both terms
406 describing thread-like structures. Within toxicology there is a long-standing tradition of
407 referring to such structures as fibers rather than filaments.

408 For some types of fibers, their aspect ratio has been found to determine toxicological
409 responses, for example in the case of asbestos and carbon nanotubes. Hence, from a
410 toxicological perspective it makes sense to distinguish between different shapes of plastic
411 debris using the neutral terminology described above. Additional information on the structure
412 (e.g., material porosity) can be included when relevant and only when it can be established
413 with certainty. For example, the descriptor 'foams' can draw unwanted parallels to styrofoam
414 even though several plastic types can be visually similar. A more neutral descriptor for this
415 type of porous materials would be 'expanded cellular plastics'.

416

417 **4.6 Criterion VI: Color**

418 Categorizing plastic debris according to color is useful to identify potential sources as well as
419 potential contaminations during sample preparation. As with shape, the color of an object
420 cannot easily be used to deduce the origin. Importantly, color information can be biased as

421 brighter colors are spotted more easily during visual inspection. In contrast, dark, transparent
422 or translucent particles may be underrepresented. In addition, discoloration can take place
423 during weathering as well as sample preparation, which should be considered in data
424 reporting and interpretation. While we do not find color to be crucial in a categorization
425 framework, it can make sense to include color as an additional descriptor. This can be the
426 case in a biological context, where depending on an organism's feeding preferences, some
427 colored plastic objects may be more or less likely to be mistaken as food.⁵³ As attributing
428 colors may be subjective, the use of a standardized color palette, such as the Pantone Color
429 Matching System, is preferable.

430

431 **4.7 Optional criterion VII: Origin**

432 The origin of plastic debris is commonly used as a classifier, especially for microplastics,
433 which are categorized in 'primary' and 'secondary' microplastics. In the predominant view,
434 'primary' refers to microplastics intentionally produced in that size range whereas 'secondary'
435 microplastics are formed in the environment through fragmentation or through wear and tear
436 of plastic-containing items, such as TWP and fibers released from textiles during use.¹¹ An
437 alternative perspective is that 'primary' also includes microplastics that are inherent by-
438 products of the use of a product ("primary sources"),^{35, 37} such as TWP. In that view,
439 secondary microplastics would originate from fragmentation during weathering, only. Since it
440 is challenging, if not impossible, to determine whether a particle has been generated by
441 fragmentation during intentional use or in the environment, we prefer to use the former
442 classification.

443 From a regulatory point of view, it is relevant to distinguish between primary and secondary
444 origin. This has consequences for risk management^{25, 54} as it may enable assigning
445 responsibilities and apply the polluter pays principle. However, from a biological perspective,
446 it does not matter if the plastic object encountered by an organism is intentionally
447 manufactured. In addition, while primary microplastics tend to be more uniform and

448 homogenous (e.g., microbeads), this is not always the case.⁵⁰ A subsequent weathering will
449 further change the appearance, rendering a clear-cut distinction between primary and
450 secondary (micro)plastics often infeasible. Because of this ambiguity, we suggest not to use
451 'origin' to categorize plastics unless the primary origin of plastic debris can be established
452 convincingly. One such case is the detection of microbeads originating from ion exchange
453 resins from a specific production site.⁵⁵

454 **5 Moving towards a workable terminology for plastic debris**

455 The research on plastics in the environment is still in its infancy. This makes it an exciting
456 and dynamic field but inevitably entails a certain scientific immaturity with regards to the
457 hypotheses, concepts, and methods applied. This is also true in terms of the terminology we
458 use. To promote consensus-building, we provide a framework for defining and further
459 categorizing 'plastic debris'. We identify three defining criteria and four classifiers that can be
460 used in such a framework. Based on this, we propose to define 'plastic debris' as objects
461 consisting of synthetic or heavily modified natural polymers as an essential ingredient
462 (criterion I) that, when present in natural environments without fulfilling an intended function,
463 are solid (II) and insoluble (III) at 20 °C. We further recommend using the criteria size (IV),
464 shape (V), color (VI), and origin (VII) to further categorize plastic debris (Table 1, Figure 2).
465 Each criterion covers aspects on which consensus is likely as well as elements which are
466 more debatable. Accordingly, the content of the framework cannot be fixed but may be
467 revised as the field evolves. Thus, we welcome critical input by the readers and encourage a
468 broader debate of this matter in the scientific community.

469 **Author contributions.** NBH, MW, and TH conceived and wrote the manuscript and
470 produced the figures. All co-authors critically discussed the ideas and concepts presented in,
471 commented on and reviewed the manuscript. All authors commit to organize and support the
472 post-publication consultation.

473 **Acknowledgments.** The German Water Chemistry Society supports the expert group
474 “Plastics in the aquatic environment,” whose members (TH, NB, MPH, MH, NI, MW) co-
475 authored the manuscript. MW has received funding from the European Union's Horizon 2020
476 Research and Innovation Programme under the Marie Skłodowska-Curie grant agreement
477 No 660306, from the German Federal Ministry for Transportation and Digital Infrastructure,
478 and the German Federal Ministry for Education and Research (02WRS1378, 01UU1603,
479 03F0789D).

480 **References**

- 481 1. Kühn, S.; Bravo Rebolledo, E. L.; van Franeker, J. A., Deleterious effects of litter on
482 marine life. In *Marine Anthropogenic Litter*, Bergmann, M.; Gutow, L.; Klages, M., Eds.
483 2015; pp 75-116.
- 484 2. Sedlak, D., Three lessons for the microplastics voyage. *Environ Sci Technol* **2017**, *51*,
485 (14), 7747-7748.
- 486 3. Blettler, M. C. M.; Abrial, E.; Khan, F. R.; Sivri, N.; Espinola, L. A., Freshwater plastic
487 pollution: Recognizing research biases and identifying knowledge gaps. *Water Res*
488 **2018**.
- 489 4. Gregory, M. R.; Andrady, A. L., Plastics in the marine environment. In *Plastics and the*
490 *Environment*, Andrady, A. L., Ed. John Wiley & Sons, Inc.: Hoboken, 2003.
- 491 5. Thompson, R. C.; Olsen, Y.; Mitchell, R. P.; Davis, A.; Rowland, S. J.; John, A. W. G.;
492 McGonigle, D.; Russell, A. E., Lost at sea: Where is all the plastic? *Science* **2004**, *304*,
493 (5672), 838-838.
- 494 6. Arthur, C.; Baker, J.; Bamford, H. *Proceedings of the international research workshop on*
495 *the occurrence, effects and fate of microplastic marine debris. Sept 9-11, 2008.*; National
496 Oceanic and Atmospheric Administration: 2009.
- 497 7. Koelmans, A. A.; Besseling, E.; Shim, W. J., Nanoplastics in the aquatic environment.
498 Critical review. In *Marine Anthropogenic Litter*, Bergmann, M.; Gutow, L.; Klages, M.,
499 Eds. 2015; pp 325-340.
- 500 8. Andrady, A. L., Microplastics in the marine environment. *Mar Pollut Bull* **2011**, *62*, (8),
501 1596-605.
- 502 9. Lambert, S.; Wagner, M., Characterisation of nanoplastics during the degradation of
503 polystyrene. *Chemosphere* **2016**, *145*, 265-8.
- 504 10. Gigault, J.; Halle, A. T.; Baudrimont, M.; Pascal, P. Y.; Gauffre, F.; Phi, T. L.; El Hadri,
505 H.; Grassl, B.; Reynaud, S., Current opinion: What is a nanoplastic? *Environ Pollut* **2018**,
506 235, 1030-1034.

- 507 11. GESAMP *Sources, fate and effects of microplastics in the marine environment: a global*
508 *assessment*; IMO/FAO/UNESCO-IOC/UNIDO/WMO/IAEA/UN/ UNEP/UNDP Joint
509 Group of Experts on the Scientific Aspects of Marine Environmental Protection: 2015.
- 510 12. Verschoor, A. J. *Towards a definition of microplastics - Considerations for the*
511 *specification of physico-chemical properties*; National Institute for Public Health and the
512 Environment: 2015; p 41.
- 513 13. Frias, J. P. G. L.; Nash, R., Microplastics: Finding a consensus on the definition. *Mar*
514 *Pollut Bull* **2019**, *138*, 145-147.
- 515 14. Yang, D.; Shi, H.; Li, L.; Li, J.; Jabeen, K.; Kolandhasamy, P., Microplastic pollution in
516 table salts from China. *Environ Sci Technol* **2015**, *49*, (22), 13622-7.
- 517 15. Su, L.; Xue, Y.; Li, L.; Yang, D.; Kolandhasamy, P.; Li, D.; Shi, H., Microplastics in Taihu
518 Lake, China. *Environ Pollut* **2016**, *216*, 711-719.
- 519 16. International Organization for Standardization Plastics - Vocabulary (ISO 472:2013).
520 <https://www.iso.org/obp/ui/#iso:std:iso:472:ed-4:v1:en> (16.09.2018),
- 521 17. Kole, P. J.; Lohr, A. J.; Van Belleghem, F.; Ragas, A. M. J., Wear and tear of tyres: A
522 stealthy source of microplastics in the environment. *Int J Environ Res Public Health*
523 **2017**, *14*, (10).
- 524 18. Wagner, S.; Huffer, T.; Klockner, P.; Wehrhahn, M.; Hofmann, T.; Reemtsma, T., Tire
525 wear particles in the aquatic environment - A review on generation, analysis, occurrence,
526 fate and effects. *Water Res* **2018**, *139*, 83-100.
- 527 19. Kramm, J.; Volker, C.; Wagner, M., Superficial or substantial: Why care about
528 microplastics in the Anthropocene? *Environ Sci Technol* **2018**, *52*, (6), 3336-3337.
- 529 20. Backhaus, T.; Wagner, M., Microplastics in the environment: Much ado about nothing? A
530 debate. *peerJ Preprints* **2018**, *6:e26507v6*.
- 531 21. Maynard, A. D., Don't define nanomaterials. *Nature* **2011**, *475*, (7354), 31-31.
- 532 22. Stamm, H., Nanomaterials should be defined. *Nature* **2011**, *476*, (7361), 399-399.
- 533 23. Rauscher, H.; Roebben, G.; Sanfeliu, A. B.; Emons, H.; Gibson, N.; Koeber, R.;
534 Linsinger, T.; Rasmussen, K.; Sintes, J. R.; Sokull-Klüttgen, B.; Stamm, H. *Towards a*

- 535 *review of the EC Recommendation for a definition of the term “nanomaterial”, Part 3:*
536 *Scientific-technical evaluation of options to clarify the definition and to facilitate its*
537 *implementation*; European Commission Joint Research Centre: 2015.
- 538 24. European Commission, Commission Recommendation of 18 October 2011 on the
539 definition of nanomaterial (2011/696/EU). In Official Journal of the European Union,
540 2011; p L 275/38.
- 541 25. Brennholt, N.; Heß, M.; Reifferscheid, G., Freshwater microplastics: Challenges for
542 regulation and management. In *Freshwater Microplastics*, Wagner, M.; Lambert, S., Eds.
543 2018; pp 239-272.
- 544 26. International Union of Pure and Applied Chemistry, *Compendium of polymer terminology*
545 *and nomenclature: IUPAC recommendations, 2008*. RSC Pub.: Cambridge, 2009; p 443.
- 546 27. Lechner, M. D.; Gehrke, K.; Nordmeier, E. H., *Makromolekulare Chemie*. Springer
547 Spektrum: Berlin, Heidelberg, 2014.
- 548 28. Ivleva, N. P.; Wiesheu, A. C.; Niessner, R., Microplastic in Aquatic Ecosystems. *Angew*
549 *Chem Int Ed Engl* **2017**, *56*, (7), 1720-1739.
- 550 29. Lambert, S.; Wagner, M., Environmental performance of bio-based and biodegradable
551 plastics: the road ahead. *Chem Soc Rev* **2017**, *46*, (22), 6855-6871.
- 552 30. Pfaendner, R., How will additives shape the future of plastics? *Polym Degrad Stabil*
553 **2006**, *91*, (9), 2249-2256.
- 554 31. Hahladakis, J. N.; Velis, C. A.; Weber, R.; Iacovidou, E.; Purnell, P., An overview of
555 chemical additives present in plastics: Migration, release, fate and environmental impact
556 during their use, disposal and recycling. *J Hazard Mater* **2018**, *344*, 179-199.
- 557 32. European Chemicals Agency, Guidance for monomers and polymers. In Guidance for
558 the implementation of REACH: Helsinki, 2012; p 26.
- 559 33. Lambourne, R.; Strivens, T. A., *Paint and surface coatings - theory and practice*.
560 Woodhead Publishing Ltd: Abington, 1999.
- 561 34. Verschoor, A.; Poorter, L. d.; Dröge, R.; Kuenen, J.; de Valk, E. *Emission of*
562 *microplastics and potential mitigation measures - Abrasive cleaning agents, paints and*

563 tyre wear; National Institute for Public Health and the Environment: Bilthoven, 2016; p
564 75.

565 35. MEPEX *Sources of microplastics-pollution to the marine environment*; Norwegian
566 Environment Agency: 2014; p 86.

567 36. Eunomia *Plastics in the marine environment*; Bristol, 2016; p 13.

568 37. MEPEX *Primary microplastic-pollution: Measures and reduction potentials in Norway*;
569 Norwegian Environment Agency: 2016; p 117.

570 38. Lassen, C.; Hansen, S. F.; Magnusson, K.; Norén, F.; Hartmann, N. I. B.; Jensen, P. R.;
571 Nielsen, T. G.; Brinch, A. *Microplastics - Occurrence, effects and sources of releases to*
572 *the environment in Denmark*; The Danish Environmental Protection Agency:
573 Copenhagen, 2015.

574 39. UNECE *Globally Harmonized System of Classification and Labelling of Chemicals*
575 *(GHS) - Fifth revised edition*; United Nations: New York and Geneva, 2013.

576 40. Rogovina, L. Z.; Vasil'ev, V. G.; Braudo, E. E., Definition of the concept of polymer gel.
577 *Polym Sci Ser C* **2008**, 50, (1), 85-92.

578 41. European Chemicals Agency, Guidance on information requirements and chemical
579 safety assessment, Part B: Hazard Assessment. In Guidance for the implementation of
580 REACH: Helsinki, 2010; p 19.

581 42. Besseling, E.; Quik, J. T. K.; Sun, M.; Koelmans, A. A., Fate of nano- and microplastic in
582 freshwater systems: A modeling study. *Environ Pollut* **2017**, 220, 540-548.

583 43. Huffer, T.; Praetorius, A.; Wagner, S.; von der Kammer, F.; Hofmann, T., Microplastic
584 exposure assessment in aquatic environments: Learning from similarities and
585 differences to engineered nanoparticles. *Environ Sci Technol* **2017**, 51, (5), 2499-2507.

586 44. Scherer, C.; Brennholt, N.; Reifferscheid, G.; Wagner, M., Feeding type and
587 development drive the ingestion of microplastics by freshwater invertebrates. *Sci Rep*
588 **2017**, 7, (1), 17006.

- 589 45. International Organization for Standardization Nanotechnologies - Vocabulary, Part 1:
590 Core terms (ISO/TS 80004-1:2015). [https://www.iso.org/obp/ui/#iso:std:iso:ts:80004:-](https://www.iso.org/obp/ui/#iso:std:iso:ts:80004:-1:ed-2:v1:en)
591 [1:ed-2:v1:en](https://www.iso.org/obp/ui/#iso:std:iso:ts:80004:-1:ed-2:v1:en) (19.09.2016),
- 592 46. Hidalgo-Ruz, V.; Gutow, L.; Thompson, R. C.; Thiel, M., Microplastics in the marine
593 environment: a review of the methods used for identification and quantification. *Environ*
594 *Sci Technol* **2012**, *46*, (6), 3060-75.
- 595 47. Zhang, K.; Xiong, X.; Hu, H.; Wu, C.; Bi, Y.; Wu, Y.; Zhou, B.; Lam, P. K.; Liu, J.,
596 Occurrence and characteristics of microplastic pollution in Xiangxi Bay of Three Gorges
597 Reservoir, China. *Environ Sci Technol* **2017**, *51*, (7), 3794-3801.
- 598 48. Lusher, A. L.; Welden, N. A.; Sobral, P.; Cole, M., Sampling, isolating and identifying
599 microplastics ingested by fish and invertebrates. *Anal Methods* **2017**, *9*, (9), 1346-1360.
- 600 49. MSFD Technical Subgroup on Marine Litter *Guidance on monitoring of marine litter in*
601 *European seas - a guidance document within the common implementation strategy for*
602 *the Marine Strategy Framework Directive*; European Commission, Joint Research
603 Centre, Institute for Environment and Sustainability: Luxembourg, 2013; p 128.
- 604 50. Fendall, L. S.; Sewell, M. A., Contributing to marine pollution by washing your face:
605 microplastics in facial cleansers. *Mar Pollut Bull* **2009**, *58*, (8), 1225-8.
- 606 51. Lei, K.; Qiao, F.; Liu, Q.; Wei, Z.; Qi, H.; Cui, S.; Yue, X.; Deng, Y.; An, L., Microplastics
607 releasing from personal care and cosmetic products in China. *Mar Pollut Bull* **2017**, *123*,
608 (1-2), 122-126.
- 609 52. Wardrop, P.; Shimeta, J.; Nugegoda, D.; Morrison, P. D.; Miranda, A.; Tang, M.; Clarke,
610 B. O., Chemical Pollutants Sorbed to Ingested Microbeads from Personal Care Products
611 Accumulate in Fish. *Environ Sci Technol* **2016**, *50*, (7), 4037-44.
- 612 53. Ory, N. C.; Sobral, P.; Ferreira, J. L.; Thiel, M., Amberstripe scad *Decapterus muroadsi*
613 (Carangidae) fish ingest blue microplastics resembling their copepod prey along the
614 coast of Rapa Nui (Easter Island) in the South Pacific subtropical gyre. *Sci Total Environ*
615 **2017**, *586*, 430-437.

- 616 54. Lambert, S.; Wagner, M., Microplastics are contaminants of emerging concern in
617 freshwater environments: An overview. In *Freshwater Microplastics*, 2018; pp 1-23.
- 618 55. Mani, T.; Blarer, P.; Storck, F. R.; Pittroff, M.; Wernicke, T.; Burkhardt-Holm, P.,
619 Repeated detection of polystyrene microbeads in the lower Rhine River. *Environ Pollut*
620 **2018**, *245*, 634-641.
- 621 56. Browne, M. A.; Galloway, T.; Thompson, R., Microplastic - an emerging contaminant of
622 potential concern? *Integr Environ Assess Manag* **2007**, *3*, (4), 559-561.
- 623 57. Moore, C. J., Synthetic polymers in the marine environment: A rapidly increasing, long-
624 term threat. *Environ Res* **2008**, *108*, (2), 131-139.
- 625 58. Ryan, P. G.; Moore, C. J.; van Franeker, J. A.; Moloney, C. L., Monitoring the
626 abundance of plastic debris in the marine environment. *Philos Trans R Soc Lond B Biol*
627 *Sci* **2009**, *364*, (1526), 1999-2012.
- 628 59. Costa, M. F.; Ivar do Sul, J. A.; Silva-Cavalcanti, J. S.; Araujo, M. C.; Spengler, A.;
629 Tourinho, P. S., On the importance of size of plastic fragments and pellets on the
630 strandline: a snapshot of a Brazilian beach. *Environ Monit Assess* **2010**, *168*, (1-4), 299-
631 304.
- 632 60. Desforges, J. P.; Galbraith, M.; Dangerfield, N.; Ross, P. S., Widespread distribution of
633 microplastics in subsurface seawater in the NE Pacific Ocean. *Mar Pollut Bull* **2014**, *79*,
634 (1-2), 94-9.
- 635 61. Wagner, M.; Scherer, C.; Alvarez-Munoz, D.; Brennholt, N.; Bourrain, X.; Buchinger, S.;
636 Fries, E.; Grosbois, C.; Klasmeier, J.; Marti, T.; Rodriguez-Mozaz, S.; Urbatzka, R.;
637 Vethaak, A. D.; Winther-Nielsen, M.; Reifferscheid, G., Microplastics in freshwater
638 ecosystems: what we know and what we need to know. *Environ Sci Eur* **2014**, *26*, (1),
639 12.
- 640 62. Andrady, A. L., *Plastics and environmental sustainability*. Wiley: Hoboken, New Jersey,
641 2015; p 324.
- 642 63. Koelmans, A. A.; Kooi, M.; Law, K. L.; van Sebille, E., All is not lost: deriving a top-down
643 mass budget of plastic at sea. *Environ Res Lett* **2017**, *12*, (11).

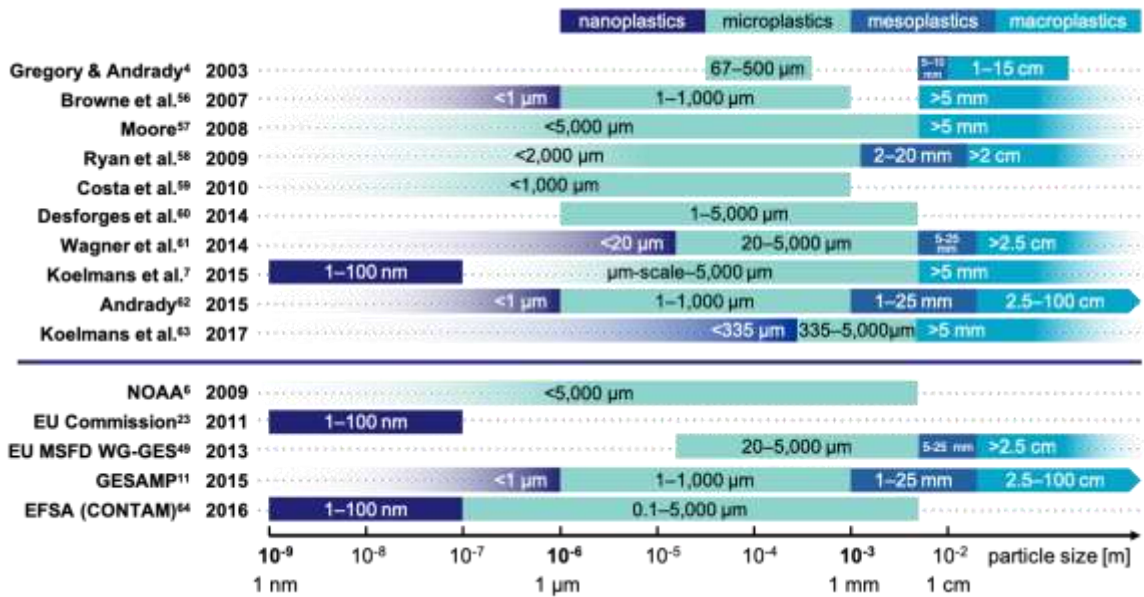
644 64. EFSA Panel on Contaminants in the Food Chain, Statement on the presence of
645 microplastics and nanoplastics in food, with particular focus on seafood. *EFSA Journal*
646 **2016**, 14, (6), 30.

648 Table 1. Overview of the recommendations for a definition and classification of plastic debris.

Criterion	Recommendation	Examples
Ia: Chemical composition		
✓ Include	All synthetic polymers: <ul style="list-style-type: none"> ▪ Thermoplastics ▪ Thermosets ▪ Elastomers ▪ Inorganic/hybrid 	All commodity plastics Polyurethanes, melamine Synthetic rubber Silicone
✓ Include	Heavily modified natural polymers (semi-synthetic)	Vulcanized natural rubber, regenerated cellulose
× Exclude	Slightly modified natural polymers	Dyed natural fibers
Ib: Additives		
✓ Include	All polymers included in Ia disregarding their additive content	Plasticized PVC with >50 % additives
Ic: Copolymers		
✓ Include	All copolymers	ABS, EVA, SBR
Id: Composites		
✓ Include	All composites containing synthetic polymer as essential ingredient	Reinforced polyester and epoxy
✓ Include	All surface coatings containing polymers as essential ingredient	Paints containing polyester, PUR, alkyd, acrylic, epoxy resin
✓ Include	Tire wear (and road) particles	-
? Open question	Is it necessary to define a minimum polymer content?	
II: Solid state		
✓ Include	All polymers with a T_m or $T_g > 20$ °C	See examples in Ia
× Exclude	Polymer gels	PVA, PEG
? Open question	Should wax-like polymers ($T_g < 20$ °C) be included?	
III: Solubility		
✓ Include	All polymers with a solubility < 1 mg L ⁻¹ at 20 °C	See examples in Ia
IV: Size		
	<ul style="list-style-type: none"> ▪ Nanoplastics: 1 to $< 1,000$ nm ▪ Microplastics: 1 to $< 1,000$ μm ▪ Mesoplastics: 1 to < 10 mm ▪ Macroplastics: 1 cm and larger <p>The largest dimension of the object determines the category. Comprehensive reporting of dimensions is preferred (e.g., for fibers).</p>	
V: Shape and structure		
	<p>Spheres: Every surface point has the same distance from the center</p> <p>Spheroid: Imperfect but approximate sphere</p> <p>Cylindrical pellet: Rod-shaped, cylindrical object</p> <p>Fragment: Particle with irregular shape</p> <p>Film: Planar, considerably smaller in one than in the other dimensions</p> <p>Fiber: Significantly longer in one than wide in two dimensions</p> <p>Additional information on the structure (e.g., porosity) can be included.</p>	
VI: Color		
	Not crucial for a categorization but useful in a biological context (e.g., when color is a cue for ingestion). Use a standardized color palette.	
VII: Origin		
	<p>Primary: Intentionally produced in a certain size</p> <p>Secondary: Formed by fragmentation in the environment or during use</p> <p>Origin should only be used if the primary origin can be established.</p>	

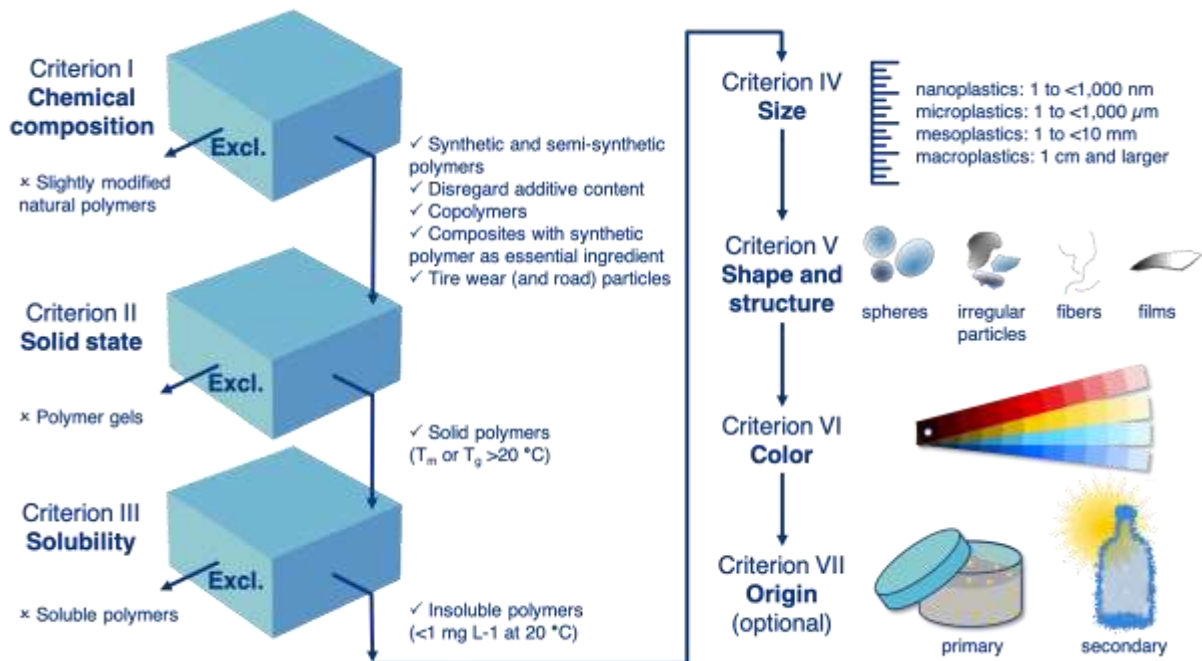
650 **Figures**

651



652

653 **Figure 1.** Examples of differences in the categorization of plastic debris according to size as
 654 applied (and/or defined) in scientific literature and in institutional reports. It should be noted
 655 that this does not represent an exhaustive overview of all used size classes.



656

657 **Figure 2.** Proposed definition and categorization framework. Excl. = excluded, see Table 1
 658 for details on criteria.